

## Effect of Mixing Individual Isolates for Bioreduction of Metals in Contaminated Soil

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**Abstract:** Bioremediation of metal contaminated soil involves complex processes, and the non-degradable nature of metals makes it more difficult. The known sources of metal pollution are many and adopting a specific remediation procedure for it is often impossible. Microbes found at polluted sites tend to serve as good remediation agents due to perceived environmental potential. Therefore, it is imperative to select bioremediation by enhancing microbes potential as a way of restoring metal polluted site to original or undisturbed state. Hence, this study tried to evaluate the individual effect of bacterial isolates on the remediation of metal polluted soil from landfill, against combined impact of microbes blended together. Soil samples that characterized of  $\text{Cd}^{2+}$ ,  $\text{Cr}^{2+}$ ,  $\text{Pb}^{2+}$  and  $\text{Zn}^{2+}$  were subjected to bioremediation. Results revealed no more than 50% reduction across the treatments amended with individual isolates. Rather higher metal remediation was observed when individual isolates were mixed prior to application. Therefore, the study suggests that individual isolates from contaminated sites can selectively metabolize to bioreduce metal concentrations. However, a mix of such isolates with manipulation in the concentration level can enhance metal reduction efficiency (>50%).

**Key words:** Bioaugmentation, bioremediation, metal pollution, leachate.

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### 1. Introduction

Environmental restoration/recovery is the core way of maintaining a functional ecosystem affected by pollution/contamination. One common method of environmental restoration is bioremediation of contaminated environments, especially soil core. Inundation of most soil cores with metals above natural occurrence levels is common, and this pose ecological risk to both humans and lower organisms, including microbes. Metal pollution of soil is common due to the recent wave of industrialization and economic development. This developmental problem is common with the developing economies of the world and other nations that are not up-to-date with the state-of-the-art technologies that minimize pollution. Prevalent in such communities is metal pollution that often emanate from indiscriminate metal waste disposal, landfilling and mine tailings. Leachate from landfills is a major pollutant fluid that contains metals, and easily permeate soil core, groundwater and even runs off into surface water. In fact, metals from leachate mobilize even kilometres away from landfill site [1]. Hence, this is the reason bioremediation of leachate-metal polluted soil is very important. A sustainable soil remediation approach is any method that involves technical removal of pollutants while soil quality and functioning are preserved [2], [3]. Basically,

bioremediation is primarily the use of degrader microbes, but the toxic level of contaminants within a polluted soil environment may only allow the survival of some microbes rather than giving chance for pollutant degradation (in case of most organic compounds) or transformation (in case of most inorganic compounds like metals) [4]. However, when microbes are properly engineered or enhanced with respect to population, biochemical properties and diversity, they tend to optimally metabolize pollutants; metals inclusive [2]. Some bacteria species possess remediation potentials for metal pollution [5], [6]. It implies that bioaugmentation of metal polluted soil with desired microbial strain is of significant interest towards restoration of metal impacted soil. Individual strains may have varying capabilities towards polluted systems, and synergizing the potentials such of microbial strains may yield optimal response to metal pollution. Therefore, this study was designed to establish the metal remediation potentials of individual bacterial strains isolated from leachate inundated environment, and similarly evaluate the efficiency of metal reduction when bacterial strains are blended together prior to application.

## 2. Materials and Methods

Leachate contaminated soil and raw leachate samples were taken from a sanitary landfill (3027.63'N, 10100.254'E) for microbial isolation and characterization, respectively, whereas uncontaminated fresh soil samples were collected from garden (307'24.15N, 101039'16.79"E) for induced contamination experiment. Soil contamination by leachate was done at laboratory condition and its treatment was monitored there. Hence, 2 kg of the soil samples were weighed into perforated plastic bags, and subsequently contaminated with leachate (10% v/w). The contaminated soil samples were allowed to stand for 48 hours for pre-diffusion of leachate into the soil system.

Subsequently, bacterial species isolated from the contaminated landfill soil were reactivated and sub-cultured using Nutrient Broth E and Nutrient Agar, respectively, in accordance to standard methods. Individual microbes, namely *Bacillus* sp, *Lysinibacillus* sp and *Rhodococcus* sp were used to prepare discrete inoculums in flasks containing Nutrient Broth E after steady growth phase was reached (1.3 ABS at OD600nm). Microcosms (4) were set-up based on amendment with individual inoculums; hence the treatments were designated A (amended with *Bacillus* sp), B (amended with *Lysinibacillus* sp), C (amended with *Rhodococcus* sp) and D (control without any amendment). Each treatment was set-up in triplicates. Another microcosm that contained mixture of inoculums of all isolated bacteria strains was prepared after monitoring the performance of the individual isolates.

About 60 – 65% soil moisture level was maintained throughout the experiment, and the bioremediation activity was monitored at 20 days intervals for the next 100 days. Some physicochemical properties of the soil were observed alongside the microbial population within the system. Samples taken on each day of monitoring were duly analyzed for metal concentration. To 0.5 g of each soil sample, HNO<sub>3</sub> and H<sub>2</sub>O<sub>2</sub> were added [7] for acid digestion using Multiwave 3000 microwave digester (Perkin-Elmer/Anton Paar). Subsequently, the elemental concentrations were measured with Optima 5300 DV (Perkin-Elmer, Massachusetts, USA). Evaluation of a procedure blank was always carried out and other experimental precautions were taken. The study used Equation 1 to evaluate the efficiency of the system on metal reduction.

$$\% \text{ of heavy metal removal} = \left( \frac{Co(x) - CF(x)}{Co(x)} \right) \times 100\% \quad (1)$$

where

Co (x) = Initial reading of heavy metal concentration at 0th day

CF(x) = Reading of heavy metal concentration at 100th day

### 3. Results and Discussion

Two of the bacteria strains isolated from the contaminated soil namely, *Bacillus* sp and *Lysinibacillus* sp belong the genera class of Bacilli, while the other, *Rhodococcus* sp is a genera of Actinobacteria. All the bacterial species belong to the  $\gamma$ -proteobacteria group and represent a core component of native bacteria community often found in soil and on phylogenetic tree [3]. Hence, there is no anticipation that utilizing such microbes in bioremediation will cause significant alteration to the normal microbial structure in the soil. Table. 1 shows the concentrations of the studied metals at the start and end of the bioremediation study.

Table 1. Metal Distribution across the Treated Soil Microcosms

Metals	Cd	Cr	Pb	Zn
Initial Conc. (mg/kg)	1.3	10	70.4	121
<b>Treatment A</b>				
a (mg/kg)	1.2	9	60.8	91
b (%)	7.6	10	13.6	25
<b>Treatment B</b>				
a (mg/kg)	1.1	8	61.8	79
b (%)	15.3	20	12.2	34.7
<b>Treatment C</b>				
a (mg/kg)	1.1	9	52.5	96
b (%)	15.3	10	25.4	20.6
<b>Treatment D</b>				
a (mg/kg)	1.3	10	6.19	101.7
b (%)	0.0	0.0	12.1	16.5

\*a –residual metal concentration; b –percentage of removal

About 35% of Zn was reduced from Treatment B. It reflected the highest percentage for Zn, and other metals reduced across all the treatments characterized of individual isolates. Despite non-addition of inoculum to Treatment D, it still indicated that 16.5% of Zn reduction. This might be due to the presence of pre-existing microbes in the leachate-induced soil, and such microbes may have the capacity to reduce or transform portion of heavy metals [8] without prior amendment. Similarly, Treatment A recorded up to 25% reduction, which may be influenced by pH distribution (Fig. 1). Krishna et al [9] also recorded 25% of Zn reduction at neutral pH (pH 7). However, the reductions appear to be in close range across the treatments, yet the variations could be significantly traced to the individual isolates used as inoculants.

Furthermore, the maximum Pb reduction was 25% recorded at Treatment C, while 15.3% was the highest reduced Cd concentration in Treatments B and C. Basically, the result showed that each treatment amended with any of the isolates yielded more metal reduction than the control experiment (no additional microbial augmentation). The degree of metal reduction across all amended treatments were less than 50% and does not reflect the optimal impact of microbes as good remediates. However, the distribution of the reduction

showed that immobilization was taking place, and can be a good indication of the potentials of the utilized bacteria species. In fact, the measured redox potential across treatments (Fig. 2) may indicate promising impact of the microbes because variation of redox potential indicates solubility of metals in contaminated soil.

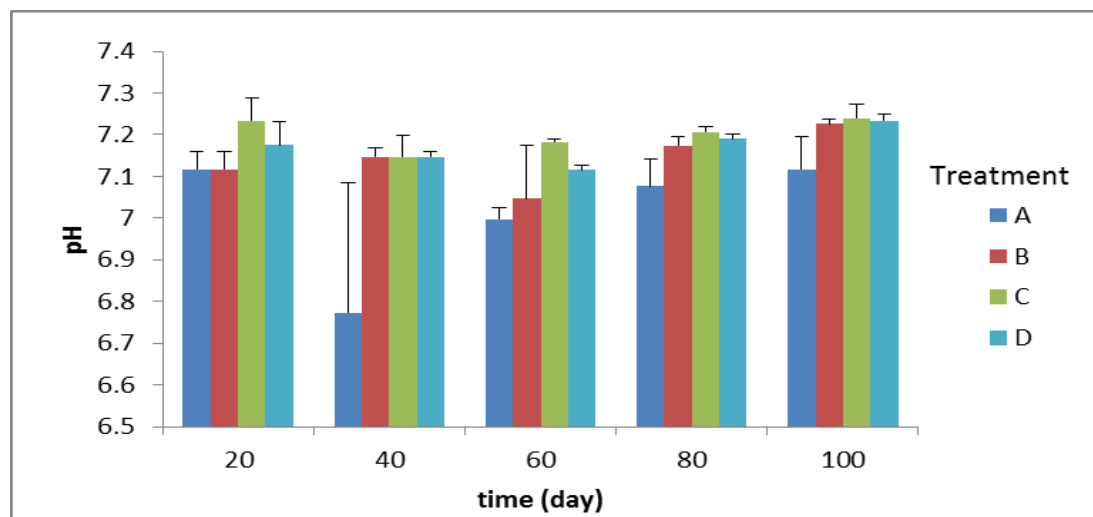


Fig. 1. pH values for each individual isolate treatment during bioremediation.

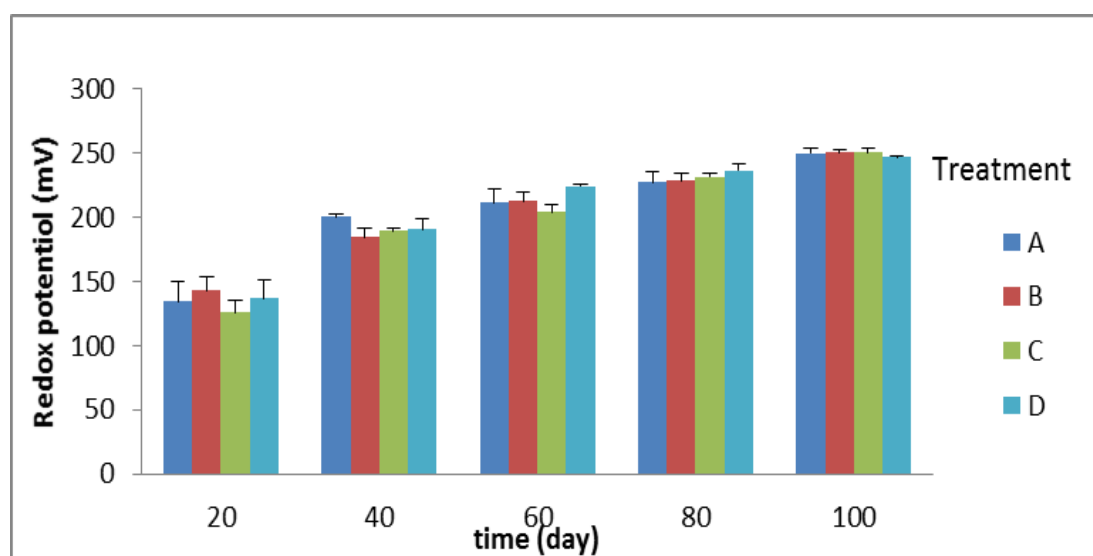


Fig. 2. Redox potential (Eh) values for each individual isolate treatment during bioremediation.

At 20th day of bioremediation, Treatment C showed the lowest value in Eh (125.56 mV), meanwhile Treatment B showed the highest with 142.9 mV. However, at final day of bioremediation, all treatments showed increase in redox potential value with no significant difference across the treatments; hence the ANOVA test, showed that  $p > 0.05$  and the F value is ( $df = 4, 70$ ) 0.058. In general, this is an indication that metal reduction was taking place because increase in Eh is a reflection of decrease in solubility of metals in contaminated soil [10], [11]. With the increased Eh, we can assume that the concentration of metals in leachate contaminated soil decreases while undergoing bioremediation.

The individual assessment of the treatments showed that Treatment B recorded the highest reduction potential for the metals. It may be easy to assume that Treatment A will be more promising than others will, considering that *Bacillus* sp has been reported to be a good remediation agent, especially *B.thuringiensis*,

*B.cereus* and *B.anthraxis* [5], [12], [13]. However, *Lysinibacillus* sp often show pronounced bioremediation impact by providing a metal binding site [14], and could be the reason why Treatment B recorded the highest reduction level than others. Therefore, mixing the three bacterial isolates at equal proportions and used on leachate-metal induced soil to evaluate the potential impact on metal reduction in contaminated soil, Fig. 3 showed that more than 50% of the metals recorded significant reduction except for Cd (41.2%). At this point Cr reduction was the highest (81%) after 100 days of monitoring. It is possible that the metal binding effect of *Lysinibacillus* sp was enhanced and better expressed when in association with *Bacillus* sp and *Rhodococcus* sp which are also gram-positive bacteria [2]. Mechanisms behind the enhanced result may also be associated with increase in microbial population within a polluted environment, but most significant is the manipulation of microbial diversity. Blending microbes that exhibit mutual interaction may give optimal metabolic impact on pollutant degradation or transformation.

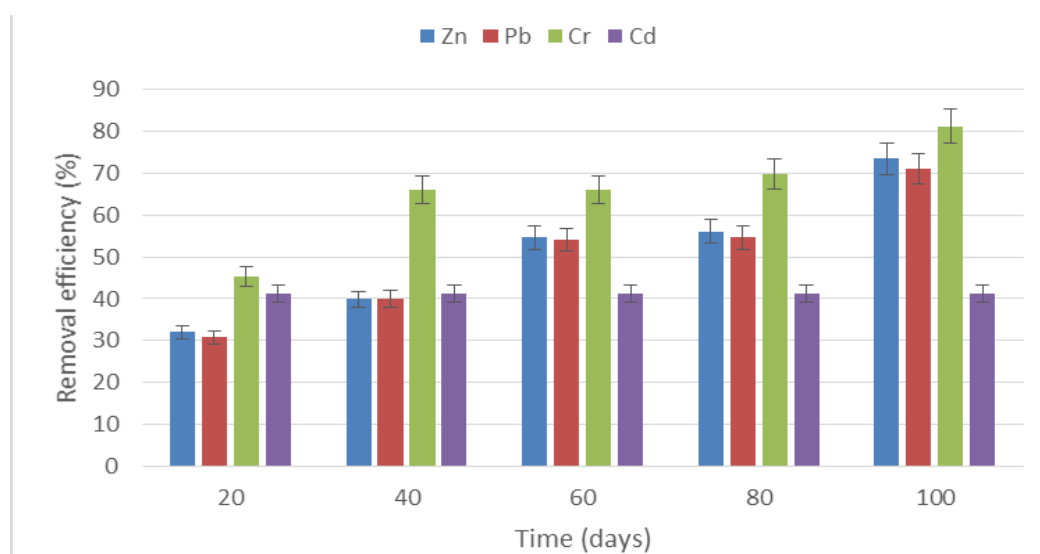


Fig. 3. Metal removal efficiency of the mixed inoculum during bioremediation.

#### 4. Conclusion

The present study deduced that leachate can induce metal pollution of soil, and the use of microbes isolated from such environment have the potential to metabolize and enhance more than 70% metal reduction. However, the microbes will not be significantly effective except the microbial population is increased. Furthermore, it was evident that the manipulation of microbial diversity allows bacteria species to perform optimally than when used as individual isolates. The individual isolates could only reduce about 8 – 35% of Cd, Cr, Pb and Zn, and does not reflect significant remediation activity. However, *Bacillus* sp when blended with *Lysinibacillus* sp and *Rhodococcus* sp can effectively reduce metal concentration in leachate-polluted soil. The reduction/immobilization efficacy of the inoculum is 70 – 80% for the identified metals except for Cd (41%).

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